

Impact of Earthquake Demolition Debris on the Quality of Groundwater

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Abstract: Problem statement: Debris from construction or demolition/deconstruction processes have no significant impact on the environment as they are res-usable and inert. This has been also long admitted for solid waste generated by the demolition of damaged cities following violent earthquakes. **Approach:** This study is a contribution to the assessment of actual impact on the quality of groundwater of buried demolition debris from the city of Boumerdes, in the North of Algeria 5 years after the May 21st 2003 earthquake hit the region. The public discharge of Boumerdes city has been used as a temporary landfill. It is located about 5 km downtown of Boumerdes at the Tidjelabine site which is marly-calcareous formation. Leachate from the landfill was directly rejected in the receiving environment, where the soil is marly-calcareous type with cracks giving a variable permeability (10^{-2} m sec⁻¹ to nearly 10^{-6} m sec⁻¹) that facilitates infiltration of potential pollutants to the groundwater. The slope character (from 5-10%) of the field contributes to pollutants movement and may accentuate water quality deterioration. Three domestic wells (designated S1, S2 and S3) were selected in the vicinity of the landfill and served as piezometers. Leachate samples were taken from the landfill and evaluated. **Results:** Leachate analysis indicated organic matter with relatively high COD (1136 mg L⁻¹ O₂) and BOD₅ (200 mg L⁻¹ O₂); whereas the pH yielded 7.65 thus indicating fermentation phase of the landfill. Heavy metal contents were beyond national standard limits except for Pb with 0.51 mg L⁻¹ which is slightly higher than limit value of 0.5 mg L⁻¹. More than five years after the creation of this landfill and despite its predominant C&D nature, these results showed that it was following a typical urban wastes decomposition scheme. Same analysis carried on water samples drawn from the piezometers yielded following results: acidic pH (6.88), acceptable values of target heavy metals concentrations except for Zn with 0.779 mg L⁻¹. Additionally bacteriological cross analysis (membrane filter and multi-tube methods) showed groundwater contamination by total coliforms (1100/100 mL), fecal coliforms (11/100 mL) and fecal streptococci (1100/100 mL). **Conclusion:** These results proved that leachate had reached the first aquifer horizon about 10 m beneath soil surface. Prior to any remediation program, Management of Boumerdes Municipality is called to quickly implement a reuse and recycling program of the demolition debris in order to stop water reservoirs contamination source.

Key words: Demolition debris, landfill, leachate, heavy metals, groundwater, infiltration

INTRODUCTION

The solid waste management is governed by standards the user must respect otherwise they expose themselves to pollution that may follows. For instance, it is common sense, that water pollution may be due to industrial effluents such as exhaust fumes and gases liquid or solid wastes that strongly contribute to water quality impoverishment. Same applies for extensive agriculture which requires fertilizers that induce increasing water pollution risks.

But, the pollution generated by solid waste from house demolition has long been underestimated as presenting no danger in the short term (Brunner and Stampfil, 1993; Tranklerlsa Walker and Dohmann, 1996).

This study is a contribution for the assessment the impact on groundwater pollution by demolition debris generated by the may 21rd, 2003 earthquake of Boumerdes. To face the emergency and urgency, demolition debris were quickly buried in temporary

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Indeed, frequent feature changes (transition from alluvial formation to cracked or compact limestone) are the reason for important variation in permeability.

As a matter of fact, we switch from a permeability of about $10^{-2} \text{ m sec}^{-1}$ to nearly $10^{-6} \text{ m sec}^{-1}$.

Thus, the flow directions follow existing cracks.

However, the hydro-geological studies conducted in the area shows that there are two aquifer horizons. The first one has a relatively short depth (maximum 10 m), the alluvial Mio-Plio-Quaternary being its bottom seat and which may be polluted by inputs from the landfill. The second one is deeply located across the valangian-Albian sandstone.

Precipitations in the area average $410.5 \text{ mm year}^{-1}$ (2005/2006) and accentuate the movement of pollutants either through infiltration or by surface runoff.

Waste characterization: The town of Boumerdes covers an area of 1800 ha occupied by inhabitants (2005).



Fig. 2: Location of the Tidjelabine landfill of the city of Boumerdes

Table 1: Tonnage estimation of the most prevalent material types in disposed wastes

Material type	Est. Mass (tons)	(%)
Concrete (including iron framework)	399140	54.30
Bricks (clay)	6200	0.85
Gypsum	12680	1.75
Paints and wall coatings	2500	0.35
Lumber	3690	0.50
Plastics*	63616	8.65
Household waste**	212055	28.85
Miscellaneous***	35342	4.80

Tidjelabine landfill; *: Mainly beverage containers, grocery and trash bags, films and durable items; **: Includes food rests, stale fruits and vegetables leaves and grass, paper, textile, glass, plastic bags, domestic appliances and other small consumer electronics; ***: Includes used vehicle parts, batteries, used oil, ash, electronics, tires, asphalt, industrial sludge, glass

Table 2: Localization and use of the selected piezometers close to the Tidjelabine discharge

Designation of taking point	Situation vis-a-vis centre (O) of the landfill	Distance (m) from O
S1	East. Well of 2 m depth domestic use piezometer.	300
S2	North in residential. Well of 8 m depth domestic use	350
S3	South West in agricultural land. Well of 8 m depth irrigation use	420

The estimated masses (tons) of various types of debris buried in the site is given in Table 1 (Yost and Halstead, 1993).

MATERIELS AND METHODS

In our study, a sampling campaign and analysis was performed on the leachate from the landfill and three control wells that serve as piezometers. The collection is made to the month of March 2007 and covered the major ions, heavy metals, nitrogen, chemical applications and biological oxygen demand (COD and BOD₅), organic matter and minerals and microbiological analysis of groundwater. Temperature, pH and conductivity were measured on site.

The three wells designated S1, S2 and S3, as indicated in Fig. 2, were selected near the discharge. Table 2 gives information concerning the status of wells from the landfill. The proximity of the wells from the landfill centre is important because they are more liable and vulnerable to all forms of pollution.

Leachate:

The composition of leachate from a landfill:

Landfill leachate is similar to complex industrial waste containing both contaminating substances: organic and inorganic. Often, the inorganic contaminants are very toxic. Thus, their composition varies depending on the nature of waste, age of discharge, the technical operating and climatic conditions. Leachate may come from either waste water or rain weather and also from the water of the aquifer (Benz *et al.*, 1997).

RESULTS

The color is the first indicator of pollution. The analyzed leachate taken downstream of the landfill has a brownish color and a faecal smell, thus influencing the quality of groundwater. Results of leachate analysis are reported in Table 3. Heavy metals concentrations are compared with values of similar landfills in Table 4. Results of physical and chemical analysis of groundwater samples are given in Table 5, while their cross bacteriological states are given in Table 6 and 7. Finally, Table 8 compares target heavy metals yields found in leachate with those in groundwater of the site.

Table 3: Results of leachate samples analysis Tidjelabine discharge

Concentrations	Sample1	Sample 2	Sample 3	Sample 4	IANOR standard
pH	7.370	7.50	7.65	6.940	6.5-8.5
DCO in mg L ⁻¹	36.400	1136.00	53.76	980.98	120.0
DBO ₅ in mg L ⁻¹	8.100	198.90	1.60	145.60	35.0
MES in mg L ⁻¹	12.000	10.00	16.00	10.000	35.0
Nitrates in mg L ⁻¹	0.300	0.10	0.30	0.200	50.0
Nitrites in mg L ⁻¹	0.016	0.01	0.02	0.002	0.1
Chlorures in mg L ⁻¹	62.400	60.98	31.19	25.520	500.0
Sulfates in mg L ⁻¹	75.000	75.00	43.00	64.000	400.0
Phosphates in mg L ⁻¹	13.000	8.00	0.39	0.740	10.0
Ammoniacal nitrogen in mg L ⁻¹	0.010	0.02	0.01	0.040	30.0
Pb in mg L ⁻¹	5.890	0.02	<0.01	0.510	0.5
Zn in mg L ⁻¹	6.700	0.16	<0.01	0.470	3.0
Cd in mg L ⁻¹	<0.010	<0.01	<0.01	2.780	0.2
Cu in mg L ⁻¹	<0.010	<0.01	<0.01	<0.010	0.5

Table 4: Comparison of the levels of heavy metals in landfill leachate

Target metal mg/l	Tiaret	El jedida	Wadi akrech	Eteffont	Tidjelabine
Zn	0.50	0.0474	0.700	0.740	6.700
Cu	-	0.1580	0.450	0.270	0.050
Ni	0.60	0.1330	0.250	0.210	6.700
Cr	0.30	0.1560	0.500	0.270	3.400

Table 5: Results of physical-chemical analysis of groundwater

Parameters	S1	S2	S3
pH	6.3700	6.5600	6.8500
T °C			
CO ₃ ²⁻ mg L ⁻¹	-	-	-
HCO ₃ ⁻ g L ⁻¹	540.5800	417.9700	488.0000
Ca ²⁺ mg L ⁻¹	144.9200	116.7120	126.9730
Mg ²⁺ mg L ⁻¹	13.4200	39.4950	0.0000
Cl ⁻ mg L ⁻¹	209.1960	327.4800	98.8540
SO ₄ ²⁻ mg L ⁻¹	38.6810	139.0870	399.1550
MES	0.3000×10 ⁻²	1.7000×10 ⁻²	0.7000×10 ⁻²
DCO	74.0000	32.0000	82.0000
BDO ₅	30.0000	20.0000	40.0000
NO ₂ ⁻ mg L ⁻¹	0.0190	0.0100	0.0000
PO ₄ ²⁻ mg L ⁻¹	0.0200	0.0400	0.1400
Metal			
Cd mg L ⁻¹	0.0070	0.0078	0.0060
Cu mg L ⁻¹	0.0140	0.0047	0.0122
Pb mg L ⁻¹	0.0220	0.0000	0.0000
Zn mg L ⁻¹	0.0790	0.0124	0.7790

Table 6: Bacteriological composition of groundwater by the method of the membrane filter

Germes	well 1	well 2	well 3
Coliforms	Presence>300	Presence>300	Presence>300
Fecal Coliformes	Presence	Presence	Presence
Fecal Streptococci	Presence	Presence	Presence

Table 7: Bacteriological composition of groundwater by the method of multiple tubes

Germes	well 1	well 2	well 3
Total Coliforms	11/100 mL	28/100 mL	1100/100 mL
Fecal Coliforms	11/100 mL	3/100 mL	7/100 mL
Fecal streptococci	9/100 mL	7/100 mL	1100/100 mL

Table 8: Yields of heavy metals in leachate Vs groundwater

Heavy metal mg L ⁻¹	Leachate	Groundwater
Cd	2.78	0.078
Cu	0.10	0.014
Pb	0.51	0.022
Zn	0.47	0.779

DISCUSSION

We notice that the Chemical Oxygen Demand (COD) in leachate exceed widely accepted standards. Indeed, it is above the average standard of Algeria which is about 120 mg L⁻¹ and reached 1136 mg L⁻¹. As for BOD₅, it varies between 135 and 200 mg L⁻¹ whereas the accepted standard is 40 mg L⁻¹, thus showing significant pollution. However, the actual concentration of BOD₅ is still higher than the values found because the medium is loaded with toxins.

The concentrations of heavy metals (cadmium, chromium, zinc and nickel) are beyond acceptable standards. The concentration of lead is at the limit of the standard. Heavy metals in leachate inhibit microbial growth.

The results of chemical characterization of raw leachate from Boumerdes landfill indicated a dual pollution:

- An organic pollution that results in a high load of COD in the leachate, in sample2 for instance, the COD is about 1136 mg L⁻¹ O₂ L⁻¹ and BOD₅ is approximately 200 mg L⁻¹ O₂ L⁻¹
- A mineral pollution that results in high concentrations of some additional heavy metals in

leachate, such as in sample1 for instance with values of 3.4 mg L^{-1} for Cr, 6.7 mg L^{-1} for Ni and 6.7 mg L^{-1} for Zn

The metal composition of the leachate from the Boumerdes landfill seems to be typical of a landfill of household dominant character. Indeed, when comparing the concentrations of same metallic elements (Cu, Cr, Ni, Zn) to those of leachate generated by other garbage dumps in Tiaret (Algeria), Rabat (Marocco) or Eteffont (France), we find that the values are higher for the Tidjelabine landfill, except for Cu.

Therefore it becomes essential to recover and treat the juice from the landfill to avoid any risk of environmental contamination by infiltration of the leachate.

Applied to the analyzed leachate from the landfill, the ratio $\text{BOD}_5 / \text{COD}$ gives values ranging from 0.11-0.25, typical of an ancient but not yet stabilized landfill and corresponding to the acid phase of anaerobic degradation (Salem *et al.*, 2008).

Applied to the analyzed leachate from the landfill, the ratio gives values ranging from 0.11-0.25, typical of an ancient but not yet stabilized landfill and corresponding to the acid phase of anaerobic degradation.

Obtained results of analysis conducted on groundwater proved that the first aquifer (maximum 10 meters depth) has already been contaminated by leachate effluents, thus confirming its state and condition of polluted non-potable water.

As a matter of fact, both piezometers S1 and S2 contain weak acid waters (pH of 6.37 and 6.56 respectively) indicating the influence of the discharge on groundwater. On the other hand, S3 has a higher pH of 6.85, but still under neutrality.

For the other parameters, mainly dissolved oxygen, (NO_3 , COD, BOD_5), levels are found prove low organic matters content in groundwater, but still are higher than the accepted standards of potable water. Pick concentrations of target metals (0.078 for Cd, 0.014 for Cu, 0.022 for Pb and 0.779 for Zn) in groundwater are higher than acceptable limits.

Two methods have been used for the evaluation of bacteriological composition of groundwater in order to insure more precise results. While the membrane filter method confirmed the presence of target germs, the multi-tubes one quantified them. Obtained results show that water wells contain important pathogens (Total Coliforms up to 1100, Fecal Coliforms up to 11 and Fecal streptococci up to 1100 per 100 mL) showing a significant bacteriological contamination of groundwater. Well3 is the most polluted due to its

location on the site. This latter is located downstream of the discharge and flows follow this direction. Furthermore, the permeability of cracks would promote the infiltration of leachate.

The temperature plays a very important role in increasing bacterial activity and evaporation of water. Indeed, temperature is a key element in the enumeration of aquifer systems. It varies depending on seasonal ambient temperature, the geological nature of the soil and the depth of the aquifer level under soil surface (Tranklerlsa Walker and Dohmann, 1996). In Tidjelabine case study temperature varies between 12 and 15°C , thus low enough to avoid micro-organisms proliferation in groundwater, so the important presence of coliforms and fecal streptococci can only be explained as the result of contamination by leachate infiltration.

CONCLUSION

This study concerned the impact of debris and rubble from the demolition of cities following an earthquake and which were considered safe and inert. In the case of the Tidjelabine landfill, the results of the analysis conducted on leachate and water proved the dangerousness of this type of debris resulting in a real double impact on environment:

- A direct impact on surface water as rainfalls are polluted by leachate runoff promoted by slope configuration of the site
- An indirect impact as groundwater is polluted by leachate infiltrations through cracks of porous soil

The concentration of target heavy metals in leachate from the landfill is evolving towards a dominant municipal solid waste discharge, although it was, originally, a landfill for demolition and construction debris assumed to be inert (Durmusoglu and Yilmaz, 2006). A first explanation of this evolution may come from the presence of organic matters in the debris coming from demolished constructions by the earthquake at first hand, and from household waste later as the status of this discharge remained open for sometime after the earthquake.

The presence of germs in the piezometers shows that water is no more potable in the vicinity of the landfill, and users have been immediately informed about it. These results contribute to the enrichment of on the ground data concerning landfilled earthquake debris behavior.

Because of higher risk of contamination of our limited water resources due to leach toxicity of the

debris (Sheridan *et al.*, 2000), we strongly recommend the Management of Boumerdes to quickly address the issue through the implementation of a two-phase program:

- Reuse and recycling of the debris as a first priority, in order to eliminate the pollution source, followed by
- Remediation of the site in the second place, using appropriate treatments for heavy metals polluted soil and groundwater

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