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Developmental Design of Anaerobic Wetland System for Mining Waste Water Treatment

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Abstract: Problem statement: Tin mining is one of the oldest industries in Malaysia that have started since 1820s. These mining activities have resulted in about 13,700 hectors of tin tailings throughout the peninsula. These tailings have created numerous environmental problems such as threat to natural reserves due to landscape changes, damage to natural drainage, pollution and destruction of natural habitats. Approach: This research provided an approach for designing a constructed wetland system for treatment of tin-contaminated wastewater from mining catchment, a system that is known to provide a more economical treatment than the conventional system. Design of wetland was mostly based on the review of scientific literature, theoretical modelling and verification of performance via a pilot system. Results: Initially, physio-chemical characteristics and concentration of heavy metals in the soil and ponds were evaluated. It was found that the soil and water quality of area is highly degraded. This study will help for the design of the wetland for wastewater treatment. The study area consists of five mined out ponds in the catchment, each pond arranged in series with a 48 h hydraulic retention time. Wetland system comprises of three compartments in series-an 'inflow' pond receiving untreated tailings water overflowing into a wetland compartment, which in turn overflows into an 'outflow' pond receiving the now treated water. Each compartment filled with approximately 50 cm depth of a mixture of the cattle manure as (25%) and municipal waste compost (75%) as substrate. Waterproof baffles in each wetland compartment serve to increase the flow path of the water, thereby increasing the potential for sulphate retention. Additionally 30 tonnes of limestone will be deposited at the far end of the wetland, to facilitate final pH adjustment if it should be required. On site a computer connected to the pumps regulates the flow of tailings water through the systems. Conclusion: The planting density chosen is based on similar research on constructed wetlands. Proposed anaerobic wetland is first of its kind introduce for mining waste water treatment in Malaysia. Good results have achieved on laboratory scale operations for this system.

Key words: Water quality, heavy metals, constructed wetland, wastewater treatment, anaerobic system, physic-chemical characteristics, municipal compost, vegetation characteristic

INTRODUCTION

Wetlands are defined as having a water table above or at the soil surface for a significant proportion of the year, which is a determining factor in their make-up of the ecosystem, an emergent vegetation characteristic of wet biotopes (often containing a large proportion of helophytes), and a soil characteristic of wet biotopes (anoxic, chemically reduced). In treatment wetlands, contaminated water flows through soil, where biological and physical reactions remove contaminants (Cvetkovic and Chow-Fraser, 2011). Traditionally, treatment wetlands have been used to remove organic

and inorganic pollutants from wastewater. So most research pertaining to pollutant removal has been concerned with the biodegredation of organic and inorganic compounds.

A wetland is a more or less engineered system, designed to enhance the interaction between vegetation, fauna, soils and microorganisms for the primary purpose of pollutant removal from agricultural wastewaters (e.g., parlour washings), runoff (e.g., field, road, farmyard) or sewage (Knight et al., 2000; Ghaly et al., 2011; Michailidis, 2006). In wetland systems, water flows vertically or/and horizontally through a porous substrate (e.g., gravel, sand) planted with

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macrophytes. Normally wetlands are composed of one or more shallow, several cells of variable depth and characteristics, (depth 30-40 cm) through the vegetation, made of submerged, emergent or floatingleaved plants, which is designed to receive and treat contaminated surface water runoff from mining lakes and ponds, in such a manner that any discharge from the wetland will not pollute the water environment (Kadlec *et al.*, 2010; Steer *et al.*, 2003).

Wetlands are attractive as an endpoint in the rehabilitation of mine wastes, such as tailings and tailings water, for two reasons. First, pollutants originating from mining activities, such as metals and sulphur, are relatively immobile when present under waterlogged conditions (Zhang et al., 2010). Second, pollutants are retained by the wetlands from water passing through the wetlands (Jitrapinate et al., 2011; Keita et al., 2009). Both characteristics are largely due to the same processes. Permanently waterlogged wetland soils are generally anaerobic, because of the relatively low diffusion rate of oxygen through water compared to air. In addition, micro-organisms present in such soils respire using terminal electron acceptors other than oxygen. Such organisms can, for example, reduce ferric iron to its ferrous form, or reduce sulphate to sulphide. The formation of highly insoluble sulphide from soluble sulphate in particular is important. Not only does that process lead to the precipitation of sulphur, but also co-precipitation of metals, including iron, zinc, lead and cadmium. Once metal sulphides have precipitated, they are stable and insoluble providing the soil remains anaerobic (Lang et al., 2008). Wetlands can therefore be used in several aspects of rehabilitation of mine wastes. First, mine tailings can be revegetated under wetland conditions, using wetland plants, and second, the quality of water originating from mining operations can be improved by passing it through wetlands, whether they are naturallyoccurring or constructed specifically for that purpose (Asadi et al., 2009).

Constructed wetlands have several advantages if properly designed (Kangrang and Chaleeraktrakoon, 2007; Reza, 2008; Al-Hamed *et al.*, 2010): they can provide high and consistent level of treatment for nutrients, pathogens and hydrocarbons, contribute to runoff and flood management if built large enough, act as long-term carbon stores, are easy to manage, require little maintenance and energy use and are cheaper than alternative methods for farm runoff disposal. They minimize odours produced by agricultural wastes, due to their dense plant cover and shallow surface flow, are aesthetically pleasing if designed in a sensible manner, bring additional value to farmland and enhance habitat and biodiversity. They can be used as contingency measures against accidental spillages, for irrigation if large enough and they reduce the need for dirty water storage, decrease land area needed for application and allow better timing of land spreading.

Constructed wetlands have some limitations: their construction requires relatively large areas in comparison with conventional treatment systems and they can be costly and in the long-term and may be reduced when pollutants enter rapidly and in large amounts, and they require a minimum of water to maintain ecosystem function (Kadlec, 2006). Moreover, the creation and mismanagement of wetlands may alter existing wetlands or local hydrology, e.g. creating a pathway between the farm and water body where it was previously inexistent, can introduce invasive species, disrupt and intoxicate plant and animal communities (Heryati *et al.*, 2011; Galbrand *et al.*, 2008).

Wetlands performance varies strongly spatially and temporally, and wetlands may act as sinks or sources of contaminants, depending on their age, location, design, wastewater characteristics, loadings, retention time, hydrological conditions, season, biological activity and management (Ahmadpour *et al.*, 2010).

Vegetation in wetlands (Phragmites australis, Typha latifolia or Scirpus spp.) has an overall positive impact on treatment efficiency: it stabilizes the surface of the wetland, reduces flow velocity and facilitates sedimentation, takes up nutrients from sediment and stores them in green parts or other organs (roots, tubers), adsorbs metals, provides fixation sites for microorganisms, conducts oxygen to sediment. conditions produces aerobic which enhance nitrification, and provides wildlife with habitat and food (Nasipuri et al., 2010; Knowles et al., 2011). Plant nutrient uptake is not the major pathway for N and P removal but can contribute 16-75% removal of total nitrogen and 12-73% removal of total phosphorus. An appropriate plant selection can improve wetland efficiency: plants should be native, perennial, highly productive for rapid nutrient uptake, produce rhizome or storage organs, and be tolerant to high pollutant loads and anaerobic conditions (Vymazal, 2009). However, dying plants and accumulation of debris might increase BOD, decrease dissolved oxygen or release nutrients and affect treatment performance. Vegetation removal can be a way to export nutrients from the wetland, but it is costly, time-consuming and may disturb wetland function and decrease efficiency (Whigham, 1999).

Heavy metals (e.g., from oil spillages, mining) may be removed or stored by sedimentation, adsorption to plants and sediment, plant uptake, biological assimilation, decomposition, chemical transformation and volatilisation, these processes being mainly influenced by temperature, pH, redox potential and availability of adsorption sites (Faulwetter *et al.*, 2008; Turner *et al.*, 2000).

The variability in the design, use and performance of wetlands, and the lack of detailed studies investigating simultaneously the hydrology, ecology and economics of individual systems justifies the necessity to explore the efficiency, limitations and sustainability of the particular design used in Malaysia until now.

Wetland plants for vegetation of mine tailings: In Malaysia the approach for revegetation of mine tailings has not yet been applied, but has been proposed for the Bestari Jaya catchment. Malaysia has a net precipitation level greater than the evapotranspiration level, therefore the supply of water for the establishment of wetlands should not pose a problem. However, characteristically, mine tailings have a low nutrient content and high concentrations of potentially toxic metals and sulphur compounds, both of which can be problematic for the successful establishment of plants. Nutrient supply to the plants can be improved by adding fertiliser. Alternatively, plants that have low nutrient re- quirements can be used. The latter solution is more attractive as it reduces the cost of the reclamation process. In addition, plants that are used for revegetation purposes can survive higher metal concentrations than plants that are not accustomed to such conditions. Brooks observed that the amphibious floating sweetgrass (Glyceria fluitans) was growing very well on tailings in a pond near the abandoned lead-zinc mine at Glendalough, Co. Wicklow (Brooks et al., 2005). This was the first time that this species was reported to grow under such conditions and a study was initiated to investigate whether the species was suitable for revegetation purposes (Spieles and Mitsch, 1999).

Filtering of metals from contaminated water passing through a 'volunteer' wetland: Wetlands can also be used for quality improvement of contaminated water (Tilley and Brown, 1998).

Biogeochemical and physical processes, as well as uptake by plants, lead to reduced concentrations of contaminants, including nitrogen, phosphorus and metals, as the water passes through the wetlands. Naturally occurring, so-called 'volunteer' wetlands, as well as constructed wetlands, can be used for the treatment of polluted water. Many studies have shown the effectiveness of such systems in reducing concentrations of contaminants in water, but the question still remains as to how the system itself and its longevity are affected by the accumulation of toxic substances (Amezaga et al., 2002). If treatment wetlands deteriorate within a relatively short period of time (10-20 years) then this approach would not be attractive for municipal and industrial purposes. Most constructed treatment wetlands are younger than fifteen years and, therefore, have not been active long enough for an accurate assessment of the impact of accumulation of pollutants on their longevity. Natural, volunteer wetlands may have been receiving pollutants for a much longer period of time. Such situations are rather rare, but one example exists at Glendalough, Co. Wicklow.

A key objective of this feasibility study was to design a treatment system that would be inexpensive in terms of both initial installation costs and long-term operating and maintenance costs.

Study area: Bestari Jaya catchment is located at 3°, 24' 40.41°" N and 101° 24' 56.23" E. It is a part of Kuala Selangor district, located in Selangor, biggest state of the country. District Kuala Selangor has three main towns namely, Mukim Batang Berjuntai, Mukim Ulu Tinggi, Mukim Tg.karang. Bestari Jaya is located in Mukim Batang Berjuntai (Ashraf et al., 2011a). Tin mining activities has ceased from last ten years, now sand mining. The catchment has total of 442 small and big mining lakes and ponds (Fig. 1). Bestari Jaya has a tropical, humid climate, with very little variations in temperature throughout the year. The average temperature of the area is 32°C during day and 23°C at night (Ashraf et al., 2011b). The Bestari Java catchment is strongly impacted by mining pollution, which affects Selangor River as mining water flows freely without treatment to the river via small connecting River Ayer Hitam.

The protection of the River Selangor is a high priority due to its high ecological value and economic importance, in particular for drinking purposes and fishing which represents a significant local source of income and employment. In order to address mining pollution in the Bestari Jaya Catchment, the construction of wetlands was suggested and promoted by the writer.



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Fig. 1: Bestari Jaya Catchment showing selected ponds (Yellow) for wetland development

MATERIALS AND METHODS

Sampling: Total 92 hectors of downstream part of the catchment were sampled, starting from north-eastern side of catchment to Sungai Ayer Hitam that meets Sungai Selangor at the Jalan Timur Tambahan road junction. Water samples were taken from two exmining ponds, at the junction of Sungai Ayer Hitam and at the junction of Sungai Selangor and soil samples were taken at the embankment of the river and ponds and the area nearby. Global Positioning System GPS was used to determine the actual coordinates of the sampling sites and to reconfirm the location of the sampling site during subsequent sampling periods. Soil and water investigation consists of ten locations, in order to determine and to provide ground information for subsequent detailed planning of the future work. For soil sampling multiple sub samples were taken from each location and then samples were homogenized into composite sample with stainless spoon and then sub sampled by spoon into each sample container to get accurate results. For ex-situ analysis, soil samples were collected from first 20 cm of the soil in polythene bags and water samples were collected 10 cm below the surface water using HDPE bottle 500 mL (Noble et al., 2011). The water samples were preserved by few drops of nitric acid (70%) and stored in an icebox and transported to laboratory for analysis.

Water Investigation: Two ponds P1 and P2 at downstream of the catchment were investigated for physio-chemical parameters and heavy metals analysis. Physio-chemical parameters were analysed by instrument Hydro lab HACH MS5 while colour of water is measured by True Colour Units (TCU). For quantitave estimation of heavy metals, samples were digested by acid digestion method (ASTM D 5198-09) and analyzed by atomic absorption spectrophotometer.

Soil Investigation: Soil physico-chemical parameters measured were soil texture, temperature, hydraulic conductivity, moisture content, soil pH, and soil grain size. Texture is determined by (Simonit *et al.*, 2005), soil temperature by soil thermometer, hydraulic conductivity by (ASTM D5084-03) method, moisture content by gravimetric method, soil pH was measured by potentiometrically and Soil grain size was measured by (ASTM D422) method. For estimation of heavy metals the samples were air dried, crushed in a mortar pestle and sieved up to 0.5mm mesh sieve and then digested by wet digestion method and analysed by a Perkin Elmer AAnalyst 800 atomic absorption spectrophotometer.

RESULTS

Water quality parameters for 15 sampling stations are shown in (Table 2). Results shows that there is variation in water quality at all sampling stations.

the Bestari Jaya wetland	unitation and estin	nated costs for
Quantity/	Estimated	
Material	duration	Cost RM
Pulverised Fuel Ash PFA	614t	10000
Cattle Manure ^a	60t	2000
Municipal Waste Compost	49t	2000
Lime stone	31t	3000
Pea gravel	11T	1000
Broken stone	10t	1000
Pipe working/building material	-	5000
Top Soil	64t	5000
Design engineering/consultation	6 months	8000
Plant hire and operation	7 weeks	20000
Total		57000PM



Table 1: Construction materials quantities and estimated costs for



Contribution of heavy metals in water

Pb, 13%

□ Co, 0%

■ Mn, 9% As, 12% **-** Fe, 3%

■ Cu, 14%

 \blacksquare Pb \blacksquare Zn \blacksquare Ni \blacksquare Co \blacksquare As \blacksquare Cu \blacksquare Fe \blacksquare Mn \blacksquare Sn

Fig. 3: Contribution of heavy metals in water



Fig. 4: Concentration of heavy metals in water

Metal concentration of water and soil are good indicators of degree of contamination. (Tabl 4) indicates physo-chemical characcteristics of soil at Bestari Jaya catchment. (Table 5), (Fig. 3 and 4) show the concentration of heavy metals in water of the area under investigation. Similarly (Table 6), (Fig. 5 and 6) show the concentration of heavy metals in soil which is realy alrming. At the sampling station WS1 are as follows; lead 38, zinc 88, nickel 2.5, cobalt 1.0, arsenic 30, copper 59, iron 06, manganese 44 and tin 85 mg L^{-1} while at sampling station WS15 concentration of heavy metals are as follows; lead 96, zinc 121 nickel 2.8, cobalt 1.8, arsenic 77, copper 80, iron 16, manganese 48 and tin 250 mgL^{-1} . Same variation trends of decrease in metal concentration are at all sampling stations from upstream to downstream.

Fig. 2: Soil profile in the study area

Water quality parameters of sampling station WS15 are; colour 9 TCU, temperature 32.51°C, pH 5, conductivity 1756 µmhos/cm, salinity 0.30 %, turbidity 0.22 NTU, dissolved oxygen 6.82 mg L^{-1} , total dissolved solids 2998 mg L^{-1} while at WS1 (Junction of Sungai Ayer Hitam + Sungai Selangor) water quality parameters are; colour 5 TCU, temperature 32.19°C, pH 6.47, conductivity 1640 µmhos/cm, salinity 0.26%, turbidity 0.12 mg L^{-1} dissolved Oxygen 6.59 mg L^{-1} , total dissolved solids 2654 mg L^{-1} . This shows variation trends at all sampling stations are from upstream to downstream.

Table 3 and 4, (Fig. 2) show the physico-chemical properties of soil. (Table 3) shows the that average contents of the soil are gravel 37.3% with diameter 3-6 mm, sand 57.20% with diameter 0.1-2 mm, silt 2.9% with diameter 0.008-0.4 mm and clay 2.46% with diameter 0.0008-0.0014 mm. (Table 4) indicates that the average moisture content of soil is 6.36% of soil, temperature 22.0°C' pH5.64 and hydraulic conductivity is 13.7 cm day⁻¹.

	,		Colour	Colour Temperature		Conductivity	Salinity	Turbidity	Dissolved	Total Dissolved
Sample No.	Location	Coordinates	TCU	C°	pH	µS/cm	%	NTU	Oxygen mg/L	Solids mg/L
WS1	Junction of Sungai	3°24' 28.04" N	5.00	32.19	6.47	1640.00	0.2700	16.00	6.83	2654.00
	Selangor + Sungai Ayer Hitam	101 ⁰ 25' 54.89" E								
WS2	Junction of Sungai	3°24' 30.96" N	7.00	32.62	6.27	1680.00	0.2800	18.00	6.65	2784.00
	Ayer Hitam + Sungai Udang	101 ⁰ 25' 54.08" E								
WS3	Sungai Ayer	3° 24' 36.54" N	7.00	32.52	6.27	1686.00	0.2800	18.00	6.70	2797.00
	Hitam water flow	101º 25' 59.96" E								
WS4	Sungai Ayer	3° 24' 56.68" N	7.00	32.51	6.29	1686.00	0.2800	18.00	6.72	2800.00
	Hitam at bank of UNISEL	101 ⁰ 25' 59.18" E								
WS5	Sungai Ayer	3° 25' 06.69" N	7.00	32.98	6.29	1688.00	0.2800	19.00	6.15	2812.00
	Hitam at bank of UNISEL	101 ⁰ 26' 08.14" E								
WS6	Junction of Pond 1	3° 25' 16.57" N	7.00	32.90	6.28	1690.00	0.2800	18.00	6.42	2864.00
	to Sungai Ayer Hitam	101º 26' 09.24" E								
WS7	Pond 1	3° 25' 13.37" N	7.00	32.75	5.78	1704.00	0.2900	19.00	6.34	2900.00
		101 ⁰ 26' 04.66" E								
WS8	Pond 1	3° 25' 15.56" N	7.00	32.50	5.2	1744.00	0.2900	20.00	6.00	2934.00
		101 ⁰ 26' 0.79" E								
WS9	Junction of Pond 1	3° 25' 13.87" N	7.00	32.44	5.32	1734.00	0.2900	19.00	6.42	2924.00
	to another Pond	101º 25' 55.35" E								
WS10	Junction of Sungai	3° 25' 20.92" N	7.00	32.28	5.41	1694.00	0.2800	18.00	6.39	2887.00
	Ayer Hitam to Pond 2	101 ⁰ 26' 12.06" E		32.32	5.34	1710.00	0.2800	20.00	6.28	2912.00
	at north-eastern boundary UNISEL									
WS11	Junction of Sungai	3° 25' 22.11" N	9.00	32.12	5.22	1724.00	0.2900	22.00	6.87	2920.00
	Ayer Hitam to Pond 2	101 ⁰ 26' 6.66" E								
WS12	Pond 2	3º 25' 22.54" N	9.00							
		101 ⁰ 26' 0.94" E								
WS13	Pond 2	3° 25' 22.05" N	9.00	32.57	5.39	1732.00	0.2900	24.00	6.45	2922.00
		101 ⁰ 25' 58.38" E								
WS14	Pond2	3° 25' 23.71" N	9.00	32.29	5.28	1738.00	0.2900	22.00	6.59	2956.00
		101 ⁰ 25' 52.42" E								
WS15	Junction of Pond 2	3º 25' 33.21" N	9.00	32.51	5.00	1756.00	0.3000	22.00	6.82	2998.00
	to another Pond	101 ⁰ 25' 51.34" E								
	Standard Deviation									
Mean	Variance (Standard Deviation)	Х	7.00	32.50	5.71	1707.00	0.28000	0.190	6.50	2870.00
		Ø	1.18	0.24	0.52	30.83	0.00700	2.130	0.25	87.26
		O^2	1.40	0.06	0.27	950.78	0.00006	4.552	0.06	7615.49

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Table 2: Physio-chemical parameters of surface water in the study area

Table 3: Grain size analysis of the soil

		Gravel Particle		Course Particle	o Medium		Sand Particle	Fine Particle			Silt	
Sample Diameter	Clay	Diameter		Diamete	r	Total	Diameter	Diameter			Particle	;
No. Location	Coordinates	mm	% age	mm	% age	mm	% age	% age	mm	% age	mm	% age
SS1Junction of Jalan Timur	3º 24' 29.80" N	4.00	37.66	1.00	32.01	0.100	25.34	57.35	0.040	2.81	0.0014	2.18
Tambahan + Sungai Selangor	101 ⁰ 25' 55.08" E											
SS2Bank of Sungai Ayer	3 ⁰ 24' 32.03" N	5.00	37.98	2.00	30.45	0.200	26.73	57.18	0.020	2.40	0.0009	2.44
Hitam + Sungai Udang	101° 25' 54.75" E											
SS3Bank of Sungai Ayer Hitam	3 ⁰ 24' 36.29" N 101 ⁰ 25' 57.34" E	3.00	35.44	1.00	29.48	0.100	27.52	57.00	0.020	3.88	0.0008	3.68
SS4South-eastern boundary of UNISEL	3 ⁰ 24' 54.73" N 101 ⁰ 26' 0.48" E	6.00	38.41	2.00	30.18	0.300	27.36	57.54	0.040	2.57	0.0016	1.48
SS5Wetlands developed by	3° 25' 09.78" N	5.00	36.98	1.00	31.24	0.300	25.57	56.81	0.010	2.88	0.0016	3.38
overflow of Pond 1	101° 25' 59.41" E								0.000		0.0040	
SS6Bank of Pond 1	3°25' 11.54" N 101° 26' 07.44" E	4.00	37.52	2.00	31.12	0.200	26.6	57.72	0.009	2.67	0.0019	2.09
SS7North-eastern boundary of UNISEL	3 ⁰ 25' 13.40" N 101 ⁰ 26' 11 64" F	3.00	37.84	2.00	30.52	0.200	25.92	56.44	0.020	3.13	0.0018	2.59
SS8Wetlands developed by	3 ⁰ 25' 50 18" N	4.00	36.85	2.00	20.02	0.200	26.34	56.26	0.008	3.08	0.0010	2.01
overflow of Pond 1	101 ⁰ 25' 56 90" E	4.00	50.05	2.00	27.72	0.200	20.54	50.20	0.000	5.70	0.0017	2.71
SS9Junction of Sungai Aver Hitam	3 ⁰ 25' 19.80" N	5.00	35.94	2.00	30.74	0.100	26.4	57.14	0.100	3.76	0.0016	3.16
with pond 1 on north-western side	101° 26' 13.07" E											
SS10Junction of Sungai Ayer Hitam	3 ⁰ 25' 22.79" N	5.00	37.72	2.00	29.12	0.100	27.86	56.98	0.300	2.81	0.0008	2.49
with pond 1 on south-western side	101° 26' 11.06" E											
SS11Wetland between Pond 1 and Pond 2	3º 25' 20.64" N	4.00	37.44	1.00	31.19	0.300	27.15	58.34	0.400	2.21	0.0008	2.01
	101 ⁰ 25' 54.37" E											
SS12Embankment of Pond 2	3 ⁰ 25' 27.52" N 101 ⁰ 25' 53.89" E	3.00	37.74	1.00	32.92	0.200	25.47	58.39	0.200	2.10	0.0009	1.77
SS13Embankment of Pond 2	3 ⁰ 25' 22.86" N 101 ⁰ 25' 51 67" F	6.00	37.54	2.00	31.44	0.400	26.13	57.57	0.200	2.56	0.0014	2.33
SS14Embankment of Pond 2	3 ⁰ 25' 34.95" N	3.00	37.75	2.00	30.88	0.300	25.71	56.59	0.300	3.12	0.0016	2.54
SS15Embankment of Pond 2	3 ^o 25' 36.24" N 101 ^o 25' 52.14" E	5.00	37.58	1.00	31.12	0.200	26.76	57.88	0.009	2.62	0.0018	1.92
Mean	Х	4.33	37.35	1.60	30.82	0.200	26.45	57.20	0.110	2.90	0.0013	2.46
Standard Deviation	Ø	1.04	0.77	0.50	0.95	0.090	0.78	0.63	0.130	0.57	0.0004	0.61
Variance (Standard Deviation)	O^2	1.09	0.60	0.25	0.91	0.008	0.61	0.40	0.010	0.33	0.0000	0.37

Sample	•		Moisture Content	(%)	Hydraulic cond	luctivity
No.	Location	Coordinates	by weight	Temperature °C	Ph	cm/day
SS1	Junction of Jalan Timur	3°24' 29.80" N	6.230	21.22	5.70	14.30
	Tambahan + Sungai Selangor	101° 25' 55.08" E				
SS2	Bank of Sungai Ayer Hitam	3°24' 32.03" N	6.240	22.84	5.50	12.40
	+ Sungai Udang	101° 25' 54.75" E				
SS3	Bank of Sungai	3°24' 36.29" N	6.380	21.44	5.30	15.20
	Ayer Hitam	101 [°] 25' 57.34" E				
SS4	South-eastern boundary	3°24' 54.73" N	6.430	22.19	5.30	12.90
	of UNISEL	101° 26' 0.48" E				
SS5	Wetlands developed by	3° 25' 09.78" N	6.520	22.92	5.40	13.80
	overflow of Pond 1	101° 25' 59.41" E				
SS6	Bank of Pond 1	3° 25' 11.54" N	6.180	21.14	5.80	12.50
		101 ⁰ 26' 07.44" E				
SS7	North-eastern	3° 25' 13.40" N	6.340	22.81	5.70	13.60
	boundary of UNISEL	101 ⁰ 26' 11.64" E				
SS8	Wetlands developed by	3° 25' 59.18" N	6.480	22.45	5.80	13.70
	overflow of Pond 1	101° 25' 56.90" E				
SS9	Junction of Sungai Ayer Hitam with	3°25' 19.80" N	6.440	21.91	5.80	14.10
	pond 1 on north-western side	101° 26' 13.07" E				
SS10	Junction of Sungai Ayer Hitam	3° 25' 22.79" N	6.300	21.65	5.80	14.50
	with pond 1 on south-western side	101° 26' 11.06" E				
SS11	Wetland between	3°25' 20.64" N	6.380	22.24	5.40	14.80
	Pond 1 and Pond 2	101 [°] 25' 54.37" E				
SS12	Embankment	3° 25' 27.52" N	6.410	21.14	5.80	13.80
	of Pond 2	101° 25' 53.89" E				
SS13	Embankment	3° 25' 22.86" N	6.280	22.34	5.10	12.90
	of Pond 2	101° 25' 51.67" E				
SS14	Embankment	3° 25' 34.95" N	6.390	21.87	5.80	14.60
	of Pond 2	101° 25' 49.93" E				
SS15	Embankment of	3°25' 36.24" N	6.430	22.39	5.70	13.30
	Pond 2	101° 25' 52.14" E				
Mean	Х	6.36	22.000	5.60	13.70	
Standard D	eviation	Ő	0.090	0.61	0.23	0.84
Variance (S	Standard Deviation)	0 ²	0.009	0.37	0.05	0.71

Table 4: Physio-chemical parameters of soil in the study area

Table 5: Heavy metals concentration in the surface water of the study area

Sampla			Element Concentration mg/l (ppm)								
No.	Location	Coordinates	Pb ²⁺	Zn ²⁺	Ni ²⁺	Co^{2+}	As ³⁺	Cu^{2+}	Fe ^{2_}	Mn ²⁺	Sn^{2+}
WS1	Junction of Sungai Selangor	3°24' 28.04" N	38.00	88.00	2.50	1.00	30.00	59.00	6.0	44.00	85.00
	+ Sungai Ayer Hitam	101° 25' 54.89" E									
WS2	Junction of Sungai Ayer	3°24' 30.96" N	46.00	86.00	2.50	2.10	35.00	78.00	10.0	46.00	100.00
	Hitam + Sungai Udang	101° 25' 54.08" E									
WS3	Sungai Ayer Hitam	3°24' 36.54" N	45.00	86.00	3.10	2.00	32.00	68.00	12.0	46.00	150.00
	water flow	101° 25' 59.96" E									
WS4	Sungai Ayer Hitam	3°24' 56.68" N	51.00	87.00	3.60	1.90	36.00	76.00	15.0	47.00	150.00
	at bank of UNISEL	101° 25' 59.18" E									
WS5	Sungai Ayer Hitam	3° 25' 06.69" N	51.00	86.00	2.90	2.00	52.00	69.00	13.0	49.00	155.00
	at bank of UNISEL	101° 26' 08.14" E									
WS6	Junction of Pond 1	3° 25' 16.57" N	60.00	88.00	7.50	2.90	78.00	71.00	10.0	49.00	200.00
	to Sungai Ayer Hitam	101° 26' 09.24" E									
WS7	Pond 1	3° 25' 13.37" N	58.00	88.00	8.10	2.50	91.00	60.00	12.0	48.00	225.00
		101° 26' 04.66" E									
WS8	Pond 1	3° 25' 15.56" N	89.00	90.00	6.20	2.80	88.00	80.00	15.0	49.00	268.00
		101° 26' 0.79" E									
WS9	Junction of Pond1 to another Pond	3°25' 13.87" N	67.00	90.00	4.30	3.00	67.00	75.00	20.0	48.00	227.00
	Pond 2 at north-eastern boundary UNISEL	101° 25' 55.35" E									
WS10	Junction of Sungai	3° 25' 20.92" N	80.00	92.00	3.40	2.70	91.00	70.00	14.0	48.00	199.00
	Ayer Hitam to	101° 26' 12.06" E									
WS11	Junction of Sungai	3° 25' 22.11" N	89.00	94.00	5.90	2.90	69.00	78.00	18.0	49.00	134.00
	Ayer Hitam to Pond 2	101° 26' 6.66" E									
WS12	Pond 2	3° 25' 22.54" N	91.00	132.00	8.10	1.80	90.00	95.00	19.0	51.00	155.00
		101° 26' 0.94" E									
WS13	Pond 2	3°25' 22.05" N	87.00	110.00	6.20	2.10	89.00	81.00	20.0	50.00	190.00
		101° 25' 58.38" E									
WS14	Pond2	3° 25' 23.71" N	94.00	122.00	5.50	2.50	71.00	88.00	18.0	49.00	198.00
		101° 25' 52.42" E									
WS15	Junction of Pond 2	3°25' 33.21" N	96.00	121.00	2.80	1.80	77.00	80.00	16.0	48.00	250.00
	to another Pond	101° 25' 51.34" E									
	Standard Deviation	-									
Mean	Variance (Standard Deviation)	Х	69.46	87.8	4.80	2.20	66.00	75.00	14.00	48.00	179.00
		0	20.7	31.96	2.06	0.55	23.36	9.56	4.10	1.75	52.53
		Ö ²	428.55	1021.70	4.24	0.30	546.11	91.45	16.83	3.06	2760.00

~			Element	Concentrati	on mg/kg						
Sampl	e Location	Coordinatas	 Dh ²⁺	7 n ²⁺	N;2+	Co^{2+}	A o ³⁺	$C u^{2+}$	$\mathbf{E}a^2$	Mn ²⁺	S m ²⁺
10.	Location	coordinates	FU	2.11	111	2.00	AS	Cu	10	NIII 04.00	311
881	Junction of Jalan Timur	3° 24′ 29.80″ N	110.00	120.00	8.50	3.00	70.0	120.00	22.0	84.00	425.00
~~~	Tambahan + Sungai Selangor	101° 25' 55.08" E									
SS2	Bank of Sungai Ayer	3°24' 32.03" N	96.00	113.00	5.50	2.80	75.0	112.00	24.0	91.00	400.00
	Hitam + Sungai Udang	101° 25' 54.75" E									
SS3	Bank of Sungai Ayer Hitam	3°24' 36.29" N	110.00	132.00	6.10	2.00	82.0	128.00	25.0	72.00	390.00
		101° 25' 57.34" E									
SS4	South-eastern	3°24' 54.73" N	115.00	110.00	6.60	2.90	86.0	135.00	25.0	84.00	350.00
	boundary of UNISEL	101° 26' 0.48" E									
SS5	Wetlands developed	3° 25' 09.78" N	120.00	122.00	7.90	2.00	62.0	140.00	25.0	89.00	355.00
	by overflow of Pond 1	101° 25' 59.41" E									
SS6	Bank of Pond 1	3° 25' 11.54" N	102.00	121.00	7.50	2.90	78.0	137.00	25.0	81.00	338.00
		101 [°] 26' 07.44" E									
SS7	North-eastern	3 ⁰ 25' 13.40" N	108.00	100.00	8.10	2.50	91.0	125.00	26.0	79.00	325.00
	boundary of UNISEL	101 ⁰ 26' 11.64" E									
SS8	Wetlands developed	3 ⁰ 25' 59.18" N	99.00	120.00	6.20	2.80	88.0	100.00	26.0	86.00	368.00
	by overflow of Pond 1	101 ⁰ 25' 56.90" E									
SS9	Junction of Sungai Ayer Hitam	3 ⁰ 25' 19.80" N	97.00	102.00	7.30	3.00	67.0	125.00	28.0	98.00	387.00
	with pond 1 on north-western side	101 ⁰ 26' 13.07" E									
SS10	Junction of Sungai Ayer Hitam	3 ⁰ 25' 22.79" N	120.00	112.00	6.40	2.70	91.0	120.00	25.0	98.00	399.00
	with pond 1 on south-western side	101 ⁰ 26' 11.06" E									
SS11	Wetland between Pond 1 and	3° 25' 20.64" N	85.00	100.00	5.90	2.90	69.0	128.00	26.0	81.00	434.00
	Pond 2	101 ⁰ 25' 54.37" E									
SS12	Embankment of Pond 2	3 ⁰ 25' 27.52" N	99.00	132.00	8.10	2.80	90.0	125.00	25.0	83.00	455.00
		101 ⁰ 25' 53.89" E									
SS13	Embankment of Pond 2	3° 25' 22.86" N	97.00	110.00	6.20	2.10	89.0	130.00	28.0	90.00	490.00
		101 ⁰ 25' 51.67" E									
SS14	Embankment of Pond 2	3 ⁰ 25' 34.95" N	110.00	122.00	5.50	3.50	71.0	128.00	24.0	81.00	498.00
		101 [°] 25' 49.93" E									
SS15	Embankment of Pond 2	3° 25' 36.24" N	110.00	121.00	5.80	2.80	77.0	130.00	29.0	86.00	450.00
	Standard Deviation	101 ⁰ 25' 52.14" E									
Mean	Variance (Standard Deviation)	X	105.00	115.00	6.70	2.70	79.0	125.00	25.50	85.00	404.00
	(	Ŏ	9.81	10.24	1.02	0.41	9.8	9,94	1.76	6.92	53,46
		$O^2$	96.31	105.00	1.05	0.14	96.2	98.98	3.12	47.98	2858.00

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Fig. 6: Concentration of heavy metals in soil

#### DISCUSSION

According to results, there is variation in water quality at all sampling stations. The variation trends at all sampling stations are from upstream to downstream. Possible factors involved in this variation may include formation of wetlands, palm oil plantation and the dilution factor of water. Acidic pH and low DO is the characteristic of peat swamp water (flowing into the catchment) and also by metal and sand mining activity. The high conductivity values represent high concentration of total dissolved solids. The main source of high TDS value is the recent sand mining activity going on in the study area. This study shows that the water quality is degraded in the area.

According to values presented in Table 3, shows that the contents of the soil represent medium textured sandy soil. Sandy soils have low clay and organic matter contents and aggregation is very weak to nonexistent. The structure is called single grained. Such kind of soil cannot retain so much water and can drain quickly. Single drained soils required frequent irrigation and fertilization for plants roots to penetrate. Table 4 represents the physio-chemical characteristics of soil which indicates that soil temperature and hydraulic conductivity is feasible for plant growth but low pH due to high cations in soil and moisture content due to sandy structure depress plant growth. Similarly metal concentration of water and soil shows variation trends of decrease in metal concentration from upstream to downstreaam.

Wetland development conceptual model: (Table 1) shows the required construction materials quantities and estimated costs for the development of Bestari Jaya wetland. The estimated costs is RM57000, which is very physible and economic for this kind of system. In designing a treatment wetland for heavy metals removal, processes within three compartments must be considered: (1) water, (2) media and (3) biota (Sonntag and Cole, 2008; Hoehn *et al.*, 2003). As outlined in (Fig. 7), water is the most essential compartment; all the processes of heavy metals removal within the other compartments are encompassed within the water compartment. Metal transport is dependent upon water movement which transports the heavy metals throughout the system.

The water causes the saturation of the media, allowing reduction of the redox value in order for treatment processes to occur within the media and related biota.

Each is dependent upon the other, with water as the most important compartment. The media compartment consists of any substrate used to contain both the water and the biota. For metals removal, the media used must act as an adsorbent for metal species.



Fig. 7: Diagram noting the three compartments of a treatment wetland: water, soil and biota



Fig. 8: View of reference Pond1. The end of inlet swale is visible in the foreground; the outlet is located on the opposite bank

The ability of the media to adsorb heavy metals is dependent upon its chemical make-up, and the species of arsenic present. Media within a treatment wetland can be selected to treat different concentrations or volumes of heavy metals. The media also sustains the biota, providing water and nutrients to both plants and microorganisms. The biota compartment consists of both plants and microorganisms. Biota can uptake heavy metals and alters the redox of the environment by several means. The biota is alsoa source of organic matter for the media used to better adsorb arsenic (Mitsch *et al.*, 2008).

Biota used within a treatment wetland can be altered to best suit the conditions present.

**Proposed design for Bestari Jaya wetland:** Wetland system was proposed on the grounds of the Bestari Jaya mine tailings.

The system comprises of three compartments in series-an 'inflow' pond receiving untreated tailings water overflowing into a wetland compartment, which in turn overflows into an 'outflow' pond receiving the now treated water (Fig. 8). Waterproof baffles in each wetland compartment serve to increase the flow path of the water, thereby increasing the potential for sulphate retention. On site a computer (ACS Pentium PC) connected to the pumps regulates the flow of tailings water through the systems. Also connected to the are four permanent industrial-grade computer electrodes (Rosemount Solu Cube® Analyser Model 2700), one situated in each of the four ponds. These facilitate continuous and simultaneous monitoring of conductivity and temperature. Data are logged into a database, every half-hour for the initial two months, thereafter every three hours, 24 h a day, and can be accessed remotely via a portable modem. This makes it possible to monitor the performance of the systems from our laboratory at the University of Malaya using pc ANYWHERE 32 software.

**Reference pond 1:** Wetland is developed, by following the treatment volume approach and comprises five ponds (referred to as P1, P2, P3, P4 and P5) lined by compacted clay and separated by shallow vegetated areas submerged in wet conditions. Reference Pond is located at 3° 26' 11.10 N, 101° 26' 20.32 E, Elevation 7 m. This pond is 0.8 km from the UNISEL and Bestari Jaya Town and was selected according to its position with reference to Selangor River and water flow from the catchment. The pond has an area of 2200m², a maximum depth of 8.5m in the centre and volume of 1500 m³ (Fig. 9). Pond is estimated to receive 105 m³ h⁻¹ (i.e., a Vt of 1840 m³) runoff water from mining area including tin tailing and sand mining water with high TDS and TSS. Wastewater from the mining ponds and runoff from tin tailings will discharge into a swale (45 m long) from two pipes (50 and 80 cm in diameter) over paving slabs to minimise erosion. Water leaves reference pond P1 runs through a long shallow vegetated area (c. 40 m long, 15 m wide) and through a series of three ponds (P2: 115 m², P3: 105 m², P4: 190 m², up to 1 m deep) separated by short (c. 20 m) shallow vegetated (grass or watercress) areas.

Flow then enters a large and deep pond (P5, c.  $2500 \text{ m}^2$ , up to 1.5 m deep, vegetated) (Fig. 10 and 11). Finally, under normal conditions, water will leave pond P5 through an inlet located on the south-east corner of the pond and flows into a ditch transferring treated water to River Ayer Hitam that ultimately fed up into river Selangor. The wetland compartments will be planted with Typha latifolia (four plants per m²) and Phragmites australis (nine plants per m²).



Fig. 9: Proposed Sketch of Reference Pond 1 (not to scale) dashed arrows represent flow in subsurface pipes and full arrows represent surface flow



Fig. 10: Proposed Structure with construction materials for Pond 1 and Pond 2

Glyceria fluitans (seven plants per m²), bulbs of Iris pseudacorus (five plants per m², rhizomes of Typha latifolia (six plants per m²) Juncus effusus (seven plants per m²), Phalaris arundinacea (nine plants per m²) and Cyperus rotuduss L. (nine plants per m²). Flow rates were set at 300-500 mL min⁻¹. These rates were adapted to fit the size of the systems based on the values given for other operational systems as described by Crites (1994). The theoretical residence time for the systems is 52 days, but the applicable value has yet to be confirmed. Here we report a select number of parameters measured, namely, volunteer species (invaders), pH, redox potential, conductivity and sulphate concentrations in water.

The site was visited on a monthly or bi-weekly basis and some parameters were monitored continuously. The pH was measured using a glass combination electrode connected to a pH meter (WTW pH90). Redox potential was measured using a platinum electrode connected to a mV meter (WTW pH90). Conductivity was measured using the industrial elecelectrodes mentioned above. For the analysis of sulphate a Dionex ion chromatogram was used. Invading flora will be identified using the standard key of the Malaysian Botanical Society.



Fig. 11: Proposed sketch of wetland ponds (not to scale) at two different angles. Dashed arrows represent t underground piped flow and full arrows represent surface flow

The wetland compartment of each system is filled with approximately 50 cm depth of a mixture of cattle manure (25%) and municipal waste compost (75%). This mixture was chosen because literature shows that it combined good permeability with optimal growth of plants. At the bottom of the inflow and outflow ponds in each system, a layer of about 25 cm of a 1:6 mixture of cattle manure and municipal waste compost was deposited to provide a substrate for the invertebrate species that spontaneously inhabit the systems. The planting density chosen was based on similar research on constructed wetlands (Yang et al., 2006). Additionally 30 tonnes of limestone are deposited at the far end of the wetland, to facilitate final pH adjustment if it should be required. (Fig. 10) shows placement of limestone at the far end of the treatment system. The system was designed such that the compost depth in the wetland would be 0.30-0.50 m. An additional 0.30 m of freeboard is allowed for accumulation of material on the substrate surface. The total area of substrate surface is 440 m². To generate additional hydraulic head a concrete wall will construct across the culvert from which the discharge emanates. Two sections of 100 mm diameter pipe were built into this wall. The first carries water underground to the influent point of the wetland, discharging into a basin from where the water is distributed across the wetland. The second section of pipe allows overflow back into the original watercourse when flow rates exceed approximately 400 litres min⁻¹. Because pollutant concentrations are lower at higher flow-rates due to dilution, and because of further dilution of the overflow water by the effluent from the wetland, the impact of this water on the receiving watercourse is minimal. The water outlet structure was originally a section of 150mm diameter plastic pipe buried into the retaining embankment. A movable 90° bend on the wetland-side of this pipe allowed the water level in the wetland to be adjusted (although typically the water level has been maintained approximately 50-100 mm above the surface of the substrate (Fig. 10 and 11).

Because the site slopes downwards slightly (away from the proposed influent point to the wetland), a central weir was incorporated in the design in order that the wetland could be constructed on two levels, the second cell being 0.4 m lower than the first cell. In this way savings were made in terms of both materials costs and land area used for the embankment.

The quantities of materials used, and the overall estimated cost for the development of wetland at Bestari Jaya is given below in the Table.

#### CONCLUSION

The preliminary result obtained from this study is alarming. The results of water quality trends clearly show that majority of water quality parameters are quite high and fall in Class III in terms of Malaysian Interim Water Quality Standards. The picture is more severe if we talk in terms of heavy metals concentration in the area. It falls above level IV in INWOS. After comparison of different parts of study area it is concluded that Bestari Jaya catchment has high pollution risks on environment, Sungai Ayer Hitam recipient of catchment water is highly polluted river that ultimately ends into Sungai Selangor, is vulnerable and sensitive ecosystem especially to metal pollution. Therefore lot of research needs to be carried out to access the pollution impact of the area on the environment and for the rehabilitation and reclamation steps to be taken.Wetlands have a strong capacity for the retention of pollutants, including those originating from mining activities. The establishment of a wetland cover over tailings provides a promising alternative to the more traditional dry land option. Critics of both applications frequently doubt the longevity of these systems. Where the use of wetlands for treatment of polluted water is concerned the answer is simple-size matters. If a wetland is built sufficiently large to manage the input of pollutants, then it should be functional for many decades (Withey and Kooten, 2011).

Restriction in the use of wetlands for treatment of wastewater is therefore determined by the available space for construction of such a system. Revegetation of tailings with wetlands should be sustainable for indefinite periods of time. The vegetation component provides the source of organic matter needed to drive the chemical reduction of sulphides and the subsequent precipitation of metal sulphides (Ningthoujam *et al.*, 2009). Through these processes the metals and sulphates are returned to the form they were derived from originally in the mining process, as many metal ores are sulphide in nature. Therefore, wetlands can be used to complete the recycling of mine wastes from sulphides back to sulphides:

- The proposed wetland at Bestari Jaya is comparatively small in engineering terms and is unique in the sense that the proposed wetland will play an important role in guiding the design of such systems for mining sites in Malaysia in future
- The wetland built is an anaerobic (compost) surface flow system. In engineering terms the decision to construct such a system was based on the limited hydraulic head available across the site

- 640 tonnes of Pulverised Fuel Ash (PFA) were used for the construction of the retaining embankments, which are keyed into the in situ soil to a depth of approximately 0.2 m. Extensive excavation of the in situ soil was not possible as it was found to be heavily contaminated with metals from former mining operations. The substrate of the wetland is a combination of horse manure, cow manure and municipal waste
- It is unclear whether temperature is the direct cause of changes in removal efficiency, or whether perhaps temperature changes influence microbial activity, which in turn affect metal removal efficiency. In either case this observation has some important implications. In particular, it suggests that wetland systems operating under cold climate conditions may be less effective, at least in terms of aluminium removal
- Three methods are currently in use for assessing constructed wetland performance
- Treatment efficiency (%)
- Area-adjustedr emoval rates (g/m2/d)
- First-order removal constants (m/d)

To make useful comparisons between constructed wetland systems a performance indicator must be independent of differences in influent pollutant concentration (Kadlec, 2006) A new method of wetland performance assessment proposed by (Kadlec, 2009) based on first-order removal of contaminants, appears to be a far better method of assessment.

**Recommendations:** Much can be drawn from the design of Bestari Jaya wetland and these lessons may be of considerable use for future constructed wetland projects:

A thorough characterisation of the quantity and quality of mine water to be treated proved essential in this project. There is no doubt that a similar familiarity should be encouraged for all such projects, since mine waters commonly exhibit fluctuations in both quantity and quality.

A key objective of the feasibility study was to design a treatment system that would be inexpensive in terms of both initial installation costs and long-term operating and maintenance costs. Investigation of the variety of construction materials available is therefore to be encouraged. Almost 50% of the total expenditure of this project is on plant hire and operation. Typically costs are incurred even when machinery is not operating due to inclement weather conditions. Establishing the exact removal mechanisms operational within the Bestari Jaya wetland will require detailed and long-term biogeochemical research, which was beyond the scope of this particular study. From the results of this work it would seem that particular emphasis needs to be placed on establishing the main mineral phases within the wetland substrate, and ascertaining the role of iron and sulphur cycling in the vicinity of the water-sediment interface.

If contaminant removal is rate dependent, as the weight of evidence suggests it is, then it is crucial to have accurate indications of residence times to properly understand the removal mechanisms operating within constructed wetlands. Tracer tests, using a conservative ion such as lithium, should be undertaken to achieve this. However, multiple tests would be required to establish residence times at different influent flow-rates. The use of automatic sampling equipment would be of great use in this regard.

The first-order removal model of assessment proposed by (Kadlec, 2006) appears to be the most appropriate method for comparing wetland performance. This being the case, future constructed wetlands may be more effectively designed on the basis of the first-order removal model. However, as (Kadlec, 2009) point out, if this is to be possible future research must be undertaken to gather values for the first-order removal constant at constructed wetlands already operational.

The anaerobic wetland treatment appear to be a very promising new treatment technology, particularly for remediation of marginally polluted mine water discharges. Previously, no research has been undertaken to determine the mineral phases accreting to the media within the reactors. Such work would certainly assist in ascertaining the exact removal mechanisms operational in these treatment units. It appears that at full-scale a very efficient water distribution system would be required for the system to operate effectively.

Wetland/passive treatment of other waste streams may be feasible, and in some cases has been successfully undertaken. Elements of the research presented here may be applicable to other water pollution issues, and an investigation of such possibilities might prove fruitful. In particular the following types of wastes may be suitable for passive treatment of landfill drainage, airport/runway drainage and sewage effluent and railway runoff.

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